

Irrigation and fertiliser strategies for minimising nitrogen leaching from turfgrass

Louise Barton and Tim Colmer

School of Plant Biology, Faculty of Natural & Agricultural Sciences, The University of Western Australia, Crawley, Western Australia, 6009. Email lbarton@agric.uwa.edu.au and tdcolmer@cyllene.uwa.edu.au

Abstract

Establishing and implementing management practices that limit N leaching from agricultural and horticultural land is a priority internationally. Movement of N through soil to surface and ground waters can degrade aquatic systems and compromise water used for drinking, industry and recreation. Reported annual rates of N leaching from turfgrass range from 0 to 160 kg N/ha/yr, representing up to 30% of applied N. Irrigation rate, fertiliser regime and turfgrass growth phase influence the amounts of N leached. Nitrogen losses tend to be low (<5% of applied fertiliser N) from established turfgrass that is not over-irrigated and has received moderate amounts of N fertiliser (i.e., 200–300 kg N/ha/yr). Efficient irrigation management is critical for efficient N use. Irrigation scheduling that does not cause water to move beyond the active rooting zone decreases the amount of N leached from established turfgrass, without being detrimental to, and in some instances enhancing, turfgrass growth and quality. Applying N fertilisers at rates and frequencies that match N requirements decreases N leaching from established turfgrass. Soil disturbance, such as during preparation of areas for planting turfgrass, can increase N leaching. Therefore the main strategies for minimising N leaching from turfgrass are i) optimise irrigation regimes, and ii) ensure N is applied at rates and frequencies that match turfgrass demand. These strategies are particularly important during turfgrass establishment. Further work is required on turfgrass-soil N cycling and partitioning of N applied to turfgrass. Research needs to be conducted for a broad range of turfgrass species, turfgrass ages, soil types and climates.

Media summary

Nitrogen leaching from turfgrass can be minimised by not over-watering and only applying fertilisers at those times of the year that the turfgrass is growing.

Key Words

Turfgrass management, thatch, preferential flow, mowing, soil amendment, soil carbon sequestration

Introduction

Nowadays people are more aware of the detrimental effects on the environment of improper use of N fertilisers. Poor N fertiliser management can cause N leaching, and increase the emissions of greenhouse and ozone depleting gases (i.e., N₂O, NO_x, NH₃). Nitrogen leaching is problematic as it can degrade surface- and ground-waters resulting in eutrophication and non-potable water supplies (OECD 1982; Smith 1998). In some countries N leaching is a significant source of water pollution (OECD 1982; Carpenter et al. 1998), and is considered difficult to control as it is often derived from extensive areas of land. Managing N leaching is also difficult because losses are often intermittent, and linked with seasonal land management activities or irregular events such as rainfall or soil disturbance. Nitrogen leaching losses will only decrease when changes in land management practices improve N use efficiency (Carpenter et al. 1998).

The contribution of turfgrass systems to N leaching is increasingly being scrutinised by communities and environmental regulators. Turfgrass generally requires regular irrigation and fertiliser applications, and is often perceived to be a source of N leaching; especially on coarse textured soils. Leaching is best minimised by ensuring N is applied at a rate that the soil-crop system is able to assimilate or utilise N (Powlson 1988; Carpenter et al. 1998). The approach taken to achieve this will vary depending on the crop N requirements, but also on the biological, chemical and physical attributes of the soil. Fertiliser N can either be taken up by turfgrass, denitrified or volatilised to gaseous N species by soil microbes, or immobilised into the soil organic matter (Petrovic 1990). Any N not involved in these processes is likely to be leached. Additional N may also become available to the crop if management practises increase net soil N mineralisation rates (Macdonald et al. 1989; Whitmore et al. 1992; Shepherd et al. 1996), and if clippings are returned to the turfgrass (Starr and DeRoo 1981; Qian et al. 2003). The ability of the

turfgrass and soil microbes to utilise applied N will also be affected by the rate that dissolved N moves through the soil profile. Plant uptake and soil biological processes often occur at greater rates in the topsoil than the subsoil. So, fertiliser and irrigation management practices that increase the contact time between applied nutrients and the topsoil should increase plant uptake and soil 'retention', and decrease N leaching. The objective of this paper is to summarise irrigation and fertiliser management strategies that minimise N leaching from turfgrass. Specifically we will report annual rates of N leaching from turfgrass, discuss the effects of irrigation, fertiliser and other turfgrass management practices on N leaching, and examine how N leaching impacts on turfgrass growth and quality.

Annual rates of nitrogen leaching from turfgrass

Soil N leaching is best quantified directly, and throughout the year so that seasonal changes in soil N availability are included in the measurement (Addiscott 1996). Measuring N leaching for an extended period will also account for any effects of establishing the experiment (e.g., soil disturbance) on N leaching. Techniques commonly used to measure N leaching from soils include porous (suction) cup lysimeters in combination with soil hydrological models, and soil lysimeters (Addiscott 1996). Although not completely without fault, these techniques are well-suited to soil types often used to grow turfgrass (i.e., sand and sandy-loam soils). For further information on techniques for measuring N leaching from soil, readers are referred to Addiscott (1990), Cameron et al. (1992), and Addiscott (1996).

Annual N leaching from turfgrass, where measurements have been taken in the field for at least 12 months using appropriate techniques, are summarised in Table 1. Only a small number of studies have reported annual N leaching rates from turfgrass, and most have been on cool-season grasses (mainly Kentucky bluegrass, *Poa pratensis*) grown in coarse (sand) to medium-textured (silt loam) soils (Petrovic 1990). Nitrogen leaching has been measured using hydrologically isolated systems (e.g., lysimeters or plots) or suction cup lysimeters in conjunction with hydrological models to predict percolation rates through the soil. Most studies have only measured inorganic N leaching losses (nitrate plus ammonium, or nitrate only), rather than total N leached. Inorganic N should not be the only form of N considered when measuring soil N leaching, as organic N has represented up to 90% of N leached in some studies (Wang and Bettany 1994; Singleton et al. 2001; Hood et al. 2003), albeit in situations not typical for turfgrasses.

Annual N leaching rates for turfgrass range from 0 to 160 kg N/ha, and represent up to 30% of the fertiliser applied N (Table 1). Irrigation rate, fertiliser regime and turfgrass growth phase all appear to affect the amount of N leached. Nitrogen losses tend to be low (i.e., <5% of applied fertiliser N) from established turfgrass that is not over-irrigated (i.e., rates equal to or less than potential evapo-transpiration) and has received moderate amounts of N fertiliser (i.e., 200–300 kg N/ha/yr). Over-irrigating and practices that disturb the soil both increase the risk of N leaching. The following sections will further discuss the effects of irrigation, fertiliser regimes, turfgrass species and soil disturbance on N leaching from turfgrass.

Turfgrass management and nitrogen leaching

Irrigation Management

Optimising irrigation management is crucial for minimising N leaching from turfgrass. Irrigation rates and frequencies that do not cause water to move beyond the active rooting zone will decrease N leaching (Brown et al. 1977; Snyder et al. 1984; Morton et al. 1988). In a two year field study, Morton et al. (1988) found that irrigation rates that caused water to percolate from the root zone of 4 year old Kentucky bluegrass increased drainage and N leaching from a sandy-loam, in comparison to field plots that were irrigated to avoid drought stress, but prevent percolation. Nitrogen leaching was equivalent to 14% of the applied N from the over-watered treatment and <3% from the scheduled irrigation treatment, which was not different to losses from plots to which no fertiliser was added. Minimising soil water movement under established Bermudagrass (*Cynodon dactylon* x *C. transvaalensis*) using a soil tensiometer-controlled irrigation system also decreased mineral-N leaching from 22% to 7.5% of applied N over 6 months (late autumn to late summer) (Snyder et al. 1984). Optimum turfgrass irrigation requirements are site-specific and will vary depending on a turfgrass species, use of the surface (e.g., turfgrass for passive areas or for sports), cultural practices (e.g., irrigation uniformity, salinity of irrigation water, mowing practices), soil type, and climate. For information on approaches to improving turfgrass irrigation efficiency, readers are referred to Carrow (2004).

Table 1. Annual nitrogen leaching losses from turfgrass

Turfgrass type & age	Soil texture & depth (mm)	Study length & method	Fertiliser type	Fertiliser rate (kg N/ha/yr)	Irrigation (mm/week)	N leached (kg N/ha/yr)	Reference
<i>Poa pratensis</i> and <i>Festuca rubra</i> , 3 years old	sandy loam 300	3 years suction cup lysimeter	inorganic ^a	0	none	0	Starr and DeRoo (1981)
			none	180–195 (2, Sp, A) ^b	none	0	
<i>Poa pratensis</i> and <i>Festuca rubra</i> , 4 years old	sandy loam 200	2 years ceramic lysimeter plates	none	0	soil sensor ^c	1.88	Morton et al. (1988)
			none	0	37.5 ^c	2.79	
			urea & Fluf [®]	97 (2, S)	soil sensor	3.04 (3.1) ^{de}	
			urea & Fluf [®]	97 (2, S)	37.5	13.65 (14) ^e	
			urea & Fluf [®]	244 (5, S, A)	soil sensor	4.87 (2.0) ^e	
		urea & Fluf [®]	244 (5, S, A)	37.5	31.94 (13) ^e		
Home lawn	sandy loam 200	2 years lysimeter plate	none	0	not given	1.3-1.4 ^e	Gold et al. (1990)
			inorganic ^a	244 (5, S, A)	not given	1.9-9.3 (<1-4) ^e	
<i>Poa pratensis</i> established from seed	silt loam 800	1 year lysimeter ^f	0	0	all treatments irrigated to prevent wilt & leaching, rate not given	38 ^e	Geron et al. (1993)
urea & RCU [§]			218 (5, Sp, S, A)	59 (27) ^e			
<i>Poa pratensis</i> established from sod			0	0		43 ^e	
			urea & RCU	218 (5, Sp, S, A)		69 (30) ^e	
<i>Poa pratensis</i> 6 years old	sandy loam 1200	2 years lysimeter	urea	196 (5, Sp, S, A)	not given	1.65 (<1) ^e	Miltner et al. (1996)
			urea	196 (5, S, A)	not given	4.05 (2.0) ^e	

See Notes on next page.

Notes to Table 1.

^a N:P:K (10:6:4) containing 50% of N as urea formaldehyde or (Years 1–2) or ammonium sulphate (Year 3) (Starr and DeRoo 1981); 50% urea and 50% urea formaldehyde (Gold et al. 1990); 50% NH₄NO₃ & 50% urea+ methyl urea (Liu et al. 1997); N:P:K (26:3:11) where N derived from urea (58%), S-coated urea (37.5%) and ammonium phosphate (4.5%) (Erickson et al. 2001). ^bNumber of fertiliser applications, and timing in parentheses, Sp = spring, S = summer, A = autumn, year = all year. ^cIrrigated mid-summer through to mid-autumn only; 37.5 mm/week represents mean maximum weekly evapotranspiration during summer months at the location of the study. ^dPercentage of N applied that leached is given in parentheses. ^eOnly inorganic-N measured in leachate. ^fStudy conducted for 1.75 years, only 1st year results shown. ^gRCU = resin coated urea. ^h*Poa pratensis*, *Festuca rubra*, *Lolium perenne*, and *Festuca longifolia*. ⁱCalculated from a mean monthly irrigation rate of 80 mm. Mean monthly evaporation for one year at the location of the study = 74 mm/mo (18.5 mm/week).

In addition to application amount, irrigation rate and frequency can also influence N leaching from turfgrass. In some soil types, the irrigation rate and frequency may cause water and N to move unevenly through the top soil via large cracks, worm holes, root channels and water repellent zones (i.e., preferential flow) (Bauters et al. 1998; McLeod et al. 2001). Preferential flow causes dissolved nutrients to move quickly through the top soil, minimising the opportunity for plant roots and soil microbes to utilise applied water and N. The irrigation rate and frequency required to minimise preferential flow will vary depending upon soil structure, and is often determined using dye or conservative tracers such as chloride or bromide (McLeod et al. 1998; McLeod et al. 2001; Nektarios et al. 2002). Generally speaking, soils that have uniformly porous structure can be irrigated at higher rates and less frequently than those soils that have large structural cracks (e.g., clays) or a tendency to develop water-repellent areas within the soil profile (e.g., ‘finger-flow’ in sands) (Starrett et al. 1995; McLeod et al. 1998; Nektarios et al. 2002). Applying surfactants to soil surfaces has been proposed as a method for minimising finger-flow in sandy soils, by mitigating localised dry spots (Wilkinson and Miller 1978) and improving water infiltration (Zartman and Bartsch 1990). However, Nektarios et al. (2002) concluded that surfactants only prevented finger-flow in a simulated United States Golf Association (USGA) putting green profile when there were no pre-existing preferential flow paths (i.e., soil irrigated frequently and never allowed to dry out). To date, decreasing irrigation rates per application, but with more frequent irrigation events, appears to be the most practical approach for preventing preferential flow in soils.

In coarse-textured soils, optimising irrigation regimes may not be sufficient to minimise N leaching. Retaining water and nutrients in the topsoil of sandy soils can be particularly difficult as these soils often have high hydraulic conductivities and low cation exchange capacities (CEC). Instead, amending these soils with inorganic or organic materials may improve soil water and nutrient retention. Engelsjord and Singh (1997) found addition of peat to sand significantly decreased the amount of N leached over 6 months from lysimeters planted with Kentucky bluegrass; with total N leaching representing 3.0 to 11.8% of applied N (8.9–35.3 kg N/ha) for the 80:20 (sand:peat) mixture and 1.8 to 3.3% of applied N (5.4–9.8 kg N/ha) for the 60:40 mixture. Increasing the proportion of peat in the sand decreased N leaching more from turfgrass fertilised monthly than from turfgrass fertilised biweekly with a soluble N fertiliser (Engelsjord and Singh 1997). Huang and Petrovic (1994) found adding clinoptilolite zeolite (9:1, w/w, soil: zeolite) decreased N leaching from creeping bentgrass by 50 to 87% when ammonium sulphate was applied to lysimeters at 196 and 293 kg N/ha. In the absence of turfgrass, the effectiveness of both inorganic and organic amendments to increase soil water holding capacity and decrease ammonium leaching from coarse textured soils has been shown to depend on the CEC of the amendment, rate of incorporation and the depth of incorporation (Adriano et al. 1980; Adriano and Weber 2001; Bigelow et al. 2001; Pathan et al. 2001; Pathan et al. 2002; Pathan et al. 2003). In cases that soil amendments have decreased N leaching from turfgrass, effects in turfgrass growth have been mixed (Huang and Petrovic 1994; Bigelow et al. 2001). While clinoptilolite zeolite increased N uptake of creeping bentgrass by up to 22% (Huang and Petrovic 1994), increasing the proportion of peat:sand decreased Kentucky bluegrass growth (5–26% depending upon fertiliser type) and colour (Engelsjord and Singh 1997). Adding relatively large quantities of materials to soils, as described in the studies cited above, presents practical challenges for incorporation to reasonable depths and can be costly.

Nitrogen Fertiliser Management

Once turfgrass irrigation management has been optimised, further decreases in N leaching may be achieved via improved N fertiliser management. Ideally, N should be applied at a rate and frequency that matches turfgrass demand, and if possible should not be applied immediately before heavy rainfall

(Brown et al. 1977; Snyder et al. 1981; Snyder et al. 1984; Morton et al. 1988; Miltner et al. 1996; Engelsjord and Singh 1997).

The amounts of N applied to established turfgrasses varies depending upon turfgrass species (Beard 1973), but typically ranges from 100 to 300 kg N/ha/yr (Petrovic 1990; Turner and Hummel 1992). At these application rates, N leaching is not significant from established turfgrass when irrigated at a rate that maintains the soil water in the rooting zone (Table 1)(Petrovic 1990). However, most turfgrass studies investigating the effects of turfgrass management on N leaching have been conducted on stands less than 15 years old (Table 1), and the effects of fertiliser application rates on N leaching from more mature stands has not been reported. It has been suggested that turfgrass N requirements decrease with time after establishment (Petrovic 1990; Qian et al. 2003). Estimates based on historical data and simulation modeling suggest that N requirements for cool-season turfgrass will be maintained for the first 10 years after establishment, and then continue to decline for up to 60 years (Petrovic 1990; Qian et al. 2003). Adjusting the fertiliser regimes to match N removal rates (plus atmospheric losses) has been proposed as an approach to minimise N leaching from older turfgrass stands (Petrovic 1990).

The amounts of N applied to turfgrasses might also need to be modified if clippings are returned. Turfgrass produces a large amount of clippings annually, representing a major pool of N. Returning clippings has been shown to reduce fertiliser N requirements by 30% (Starr and DeRoo 1981), 50% (Heckman et al. 2000) and 75% (Kopp and Guillard 2002). The effects of clipping management on N leaching have not been studied in the field. However, model simulations found that N leaching was low for 20 to 30 years after Kentucky bluegrass establishment under low N inputs (75 kg N/ha/yr) and when clippings were returned (Qian and Follett 2002). Returning clippings in combination with moderately high N inputs (150 kg N/ha/yr) increases N leaching, with predicted losses reaching 50 to 60 kg N/ha/yr by the time the turfgrass is 100 years old.

The frequency (or timing) of fertiliser applications needed to achieve quality turfgrass and minimise N leaching varies with fertiliser type. For more water-soluble N fertilisers, lower rates and more frequent applications should be used to minimise N leaching. For example, applying ammonium nitrate weekly, rather than bi-monthly, decreased N leaching from 17% to 2.5% of the N applied to established Bermudagrass irrigated using soil moisture sensors (Snyder et al. 1984). Less water-soluble fertilisers, such as slow- or control-release fertilisers can be applied at higher, less frequent rates than water-soluble fertilisers, without increasing N leaching (Snyder et al. 1984; Geron et al. 1993; Engelsjord and Singh 1997). Bermudagrass fertilised bi-monthly with a slow-release fertiliser (sulfur-coated urea) leached a similar amount of N (<5% of N applied) as turfgrass fertilised with ammonium nitrate applied weekly via fertigation (Snyder et al. 1984). Similarly, Kentucky bluegrass fertilised with a single application of sulfur-coated urea in spring leached a similar amount of N (<3% of N applied) to that fertilised bi-weekly with soluble NPK when both treatments were irrigated at a rate that prevented wilt and percolation (Engelsjord and Singh 1997). Nitrogen losses should be low from all fertiliser types as long as the N provided matches turfgrass requirements, and irrigation does not cause nutrients to move beyond the rooting zone.

The timing of fertiliser applications required to minimise N leaching also varies between 'warm'-season and 'cool'-season grasses. Applying fertilisers at times when the turfgrass is actively growing will minimise N leaching. Warm-season grasses grow best during the warmer months and at temperatures ranging from 26 to 35°C, while cool-season grasses have an optimum temperature range of 16 to 24°C (Hartley 1950; Beard 1973; Hull 1992). Consequently, it is recommended that fertilisers should be applied to established warm-season grasses from late spring through to early autumn, when growth is greatest (Turner and Hummel 1992). Applying fertiliser to warm-season grasses at cooler times of the year can increase N leaching (Brown et al. 1984; Snyder et al. 1984). For cool-season grasses, fertilising in spring and autumn is recommended (Turner and Hummel 1992) and generally results in low leaching losses under irrigation regimes designed to minimise percolation (Morton et al. 1988; Miltner et al. 1996). In some parts of the world, fertilisers are applied to cool-season grasses in late autumn, as this enhances turfgrass growth and colour the following spring (Turner and Hummel 1992). There is concern that this practice may increase the risk of N leaching during winter. However, Miltner et al. (1996) found applying fertiliser to a six year old Kentucky bluegrass in late autumn resulted in low N losses (<0.2% of applied ¹⁵N) that were similar to those after a spring application. After two years, 20% of ¹⁵N applied in

spring was recovered from the thatch, 35% from harvested clippings and 20% from the soil. For the late autumn application, 35% of ^{15}N was recovered in the thatch, 38% in the clippings and 25% from the soil. Miltner et al. (1996) attributed the low leachate losses from both spring and autumn applications to the rapid N uptake by the turfgrass (78-109% within first 18 days). Furthermore, accumulation of applied ^{15}N in the thatch after the autumn application reduced the risk of N leaching the following winter. The authors concluded fertiliser applied in late autumn to established Kentucky bluegrass does not increase the risk of N leaching when the turfgrass density and soil organic matter content is high. However, these findings do indicate that if the turfgrass density is not high and if thatch is not present (e.g., during turfgrass establishment, or periods after harvesting sod) the seasonal timing of fertiliser applications can influence N leaching (see section below 'Turfgrass management and soil nitrogen storage').

Plant testing has been suggested as an approach for better synchronizing N fertiliser applications with turfgrass N uptake (Sanchez and Doerge 1999). Traditionally, total N concentration in plant material has been used to determine if a plant contains adequate or deficient amounts for growth; however, plant sap 'quick tests' have also been developed to provide a more rapid assessment of plant N status (Handson and Shelley 1993; Lewis et al. 1993; Smith and Loneragan 1997). A basic requirement for interpretation of plant and sap analyses is the availability of reliable critical values. Critical values are often based on experimental findings or surveys of 'healthy' plants that have compared the total N concentrations in selected plant tissues with some yield parameter (Lewis et al. 1993; Reuter and Robinson 1997; Smith and Loneragan 1997). However, the development and use of critical values can be hindered by a wide range of factors including the plant part analysed, plant age, seasonal trends, nutrient interactions, and genotype (Lewis et al. 1993). Lewis et al. (1993) consider changes in critical nutrient concentrations with plant age and plant part to be one of the biggest problems limiting interpretation of plant analyses, which is particularly relevant to turfgrass management given that turfgrass stands can vary widely in age. The effectiveness of plant tissue testing as an approach for improving the timing of fertiliser applications to turfgrass and decreasing N leaching appears to be poorly documented, despite the development of critical values for total N concentrations in shoots of turfgrass species (Turner and Hummel 1992; Reuter and Robinson 1997), and the availability of sap tests.

Nitrification inhibitors have been proposed as an approach to improving fertiliser use efficiency, and thus minimise N leaching from soil (Amberger 1989). By decreasing nitrification, soil and fertiliser N is retained in the ammonium form for longer, reducing the risk of nitrate leaching. Both nitrate and ammonium can be absorbed by roots, so N maintained in the ammonium form is available for plant uptake. A number of nitrification inhibitors are available; however, the literature mainly contains information on the use of dicyandiamide (DCD) on turfgrass. Dicyandiamide has been shown to have either no benefit (Spangenberg et al. 1986; Fox and Bandel 1989; Waddington et al. 1989), or only a short-term benefit (Mosdell et al. 1986) to turfgrass growth or colour when added with urea or ammonium based fertilisers. Unfortunately, neither the effect of DCD on N leaching from fertiliser applied to turfgrass, nor the effectiveness of other nitrification inhibitors for use on turfgrass, have been reported. Ultimately, nitrification inhibitors can only be expected to improve fertiliser use efficiency if increased plant available N coincides with plant demand. Similar gains may be achieved by improving irrigation management and timing of fertiliser applications.

Choice of Turfgrass Species

Turfgrass species has been shown to directly influence nitrate leaching in a limited number of studies and generally under conditions designed to maximise leaching potential. For cool-season turfgrass species, Liu et al. (1997) ranked nitrate leaching from soil as Kentucky bluegrass > perennial ryegrass (*Lolium perenne*) > tall fescue (*Festuca arundinacea*), after measuring field losses from a silt loam for 26 months. Nitrate leaching also varied amongst cultivars within a species, with the amounts leached ranging from 3 to 45 kg $\text{NO}_3\text{-N/ha}$ for ten Kentucky bluegrass cultivars, 2 to 22 kg $\text{NO}_3\text{-N/ha}$ for ten perennial ryegrass cultivars, and 0.6 to 5 kg $\text{NO}_3\text{-N/ha}$ for ten tall fescue cultivars (Liu et al. 1997). In a study comparing nitrate leaching from under six warm-season turfgrasses, nitrate leaching was greatest for Meyer' zoysiagrass (*Zoysia japonica*; 55 kg $\text{NO}_3\text{-N/ha/yr}$) and lowest from St Augustinegrass (*Stenotaphrum secundatum*; 3 kg $\text{NO}_3\text{-N/ha/yr}$) (Bowman et al. 2002). Differences between these two warm-season grasses were attributed to differences in root length density at soil depths >300 mm, with greater root length densities improving N uptake. Root length density, rather than plant N-use efficiency (NUE), also explained differences in nitrate leaching under two genotypes of creeping Bentgrass (Bowman et al.

1998). Under high leaching conditions, 38% of applied N leached under a shallow-rooting genotype in comparison to 18% under a deeper-rooting genotype. Losses from the shallow-rooting genotype were partly mitigated by lowering the irrigation rate and delaying irrigation after applications of fertiliser (Bowman et al. 1998).

Choosing turfgrass species that require less N per unit of biomass produced may be another approach to decreasing the amounts of fertilizer applied and thus potential for leaching of N. Nitrogen-use efficiency has been shown to vary amongst turfgrass species (Liu et al. 1993) and cultivars (Jiang and Hull 1998; Jiang et al. 2000b). The physiological and metabolic basis for differences in NUE of turfgrass has not been extensively studied. For Kentucky bluegrass cultivars, NUE was negatively correlated with root-zone nitrate levels, nitrate uptake rate and nitrate reductase activity (Jiang and Hull 1998; Jiang et al. 2000b).

Turfgrass management and soil nitrogen storage

The majority of soil N is present in organic matter, in forms unavailable for plant uptake. Although organic N can be mineralised to plant available forms (e.g., NO_3^- , NH_4^+), the amount generally represents a small proportion of the organic N pool. Soil N mineralisation and immobilisation rates are influenced by soil C contents. Turfgrass management practices that increase soil C sequestration would be expected to increase soil N storage, and reduce the risk of soil N leaching. Studies of soil C cycling in turfgrass systems are scant, as long-term field studies are required (i.e., years and decades) to fully evaluate the effects of management regimes on soil C and N dynamics. Historic soil testing data suggests that cool-season turfgrass systems can sequester significant amounts of C (i.e., 1 t/ha/yr) and for up to 45 years after establishment (Qian and Follett 2002). The amount of C sequestration depends on the intensity of use of the surface, N application rate, clipping management, past land use, soil pH and soil CEC (Qian and Follett 2002; Qian et al. 2003). Model simulations predict that while C sequestration continues in a soil under turfgrass, it also continues to be a sink for applied N. However, as the rate of soil C accumulation decreases with time, the rate of soil N storage also decreases, increasing the risk of N leaching if fertiliser N applications are not adjusted (Qian et al. 2003).

Turfgrass management practices that decrease soil C sequestration would be expected to increase N mineralisation and N leaching potential. Soil C storage may decline if the turfgrass system is physically disturbed (Macdonald et al. 1989; Whitmore et al. 1992; Shepherd et al. 1996), such as during re-establishment of turfgrass areas. Establishing turfgrass has been shown to increase N leaching for up to seven months (Geron et al. 1993; Engelsjord and Singh 1997). Geron et al. (1993) measured N leaching from establishing Kentucky bluegrass for 95 weeks, and found 74 to 84% (66–83 kg NO_3^- -N/ha) of nitrate leaching occurred during the first 39 weeks following planting as seed or sod. The untreated seed control and untreated sod control also leached up to 29 to 45 kg NO_3^- -N/ha over the same period. The authors attributed greater N leaching during turfgrass establishment to increased soil mineralisation rates resulting from soil disturbance during planting. However, as N leaching was also greater from lysimeters with fertiliser applied when compared to those without fertiliser throughout the study, part of the increased loss may also have been due to applying N fertiliser. Soil C and N cycling may also change after the sudden death of turfgrass; for example, resulting from prolonged drought, herbicide misapplication, inappropriate renovation practices, and pests or diseases. Jiang et al. (2000a) reported almost 161 kg NO_3^- -N/ha/yr leached from 12 year old cool-season turfgrass plots killed using glyphosphate in early spring. In comparison, healthy plots (without fertiliser) leached 50 kg N/ha/yr during the same period. Increased nitrate leaching from the killed plots was attributed to the absence of turfgrass to utilise mineralised N, rather than enhanced mineralisation rates. Furthermore, nitrate leaching following the death of turfgrass can be partly mitigated by prompt re-seeding of such sites (Bushoven et al. 2000). In turfgrass systems soil N mineralisation may be an additional source of N during turfgrass establishment, and if mineralised N plus fertiliser N exceeds turfgrass requirements, then high N leaching losses can occur.

Knowledge on the fate of N applied to turfgrasses, and the influence of turfgrass management on soil N and C cycling is limited. Thatch, clippings and soil organic matter are known to be significant sinks for N in turfgrass systems (Petrovic 1990). However, the mechanism(s) by which thatch retains N and influences soil N cycling is not clear. In particular, what management factors influence how much N is retained by the thatch, and does N stored in thatch become a source of N leaching when a turfgrass dies or is renovated? Returning clippings to the turfgrass system provides an additional source of N that needs to

be accounted for when designing fertiliser management regimes. Approaches for managing clipping N need to be developed from an understanding of the effects of clipping management on soil N and C cycling. Information on the factors that influence soil N storage under turfgrass is also needed. Research suggests soil C sequestration can influence the magnitude of soil N storage under turfgrass (Qian et al. 2003). Qian et al (2003) recommends further studies investigating the effects on soil C sequestration on fertiliser regimes, mowing height and frequency, species, irrigation management, climate, and soil type.

Nitrogen leaching, turfgrass growth and quality

Recommended irrigation and fertiliser management strategies for minimising N leaching are unlikely to be adopted by turfgrass managers, unless it is also demonstrated that these practices do not have a detrimental effect on turfgrass growth and quality. Turfgrass managers are required to produce a surface with good cover, colour and strength; although not necessarily high biomass production, as disposal of mowing clippings and thatch is an additional expense. While it is intuitive that turfgrass management practices that reduce N leaching and maintain N within the rooting zone will also benefit turfgrass quality, the effects of irrigation and fertiliser management on both N leaching and turfgrass growth and quality are rarely reported in the same study. Notably, Snyder et al. (1984) demonstrated that optimising irrigation regimes benefited turfgrass quality and minimised N leaching. Soil-sensor controlled irrigation improved N uptake, growth and colour of Bermudagrass fertilised bi-monthly with a water-soluble fertiliser; and at the same time nitrate leaching was decreased. Matching fertiliser application frequencies to seasonal turfgrass requirements has also been shown to improve the consistency of turfgrass growth and colour, when irrigation regimes used minimised percolation (Snyder et al. 1984; Engelsjord and Singh 1997). Although single, large applications of water-soluble fertilisers initially tend to give higher turfgrass growth and colour than lower less frequent applications, these results are often short-lived and increase N leaching. Using slow-release fertilisers, or water soluble fertilisers more sparingly and frequently, appears to provide more consistent growth and colour for both warm- and cool-season grasses while minimising N leaching (Snyder et al. 1981; Snyder et al. 1984; Engelsjord and Singh 1997).

Future work

Much remains to be learnt about plant-soil N cycling under turfgrasses. Our understanding of N leaching is based mainly on data from relatively young plantings of cool-season turfgrasses receiving inorganic fertilisers (Table 1) (Petrovic 1990), and would benefit from longer-term studies of a range of species, ages, management regimes, climates and soils. Studies during the establishment phase are also needed. Turfgrass growth and quality, in addition to N leaching, should be evaluated in all future work.

Optimised irrigation regimes are crucial to control of N leaching, and technological advances have improved irrigation management (Carrow 2004). However, much of the literature on N leaching in turfgrasses has not expressed irrigation applied relative to evaporative demand, so that comparisons between sites and studies are difficult. Approaches for maintaining soil water in the rooting zone are also needed for situations where irrigation management can not be optimal, due to site (e.g., very free-draining soils, slopes, preferential flows), climate (e.g., locations prone to sudden, heavy rainfall) or cost constraints. Use of amendments to improve soil water-holding capacity, and using species that require less N, should be explored.

Quantifying N leaching is expensive and requires intensive sampling strategies over extended periods. Simulation models that predict N leaching and turfgrass performance from easily measurable parameters would assist with designing strategies for minimising N leaching from turfgrass. The success of simulation models will depend upon understanding of the fate of applied N, and the effects of management regimes on soil N and C cycling. Thatch, clippings and soil organic matter are significant sinks for N in turfgrass systems (Petrovic 1990). However, the dynamics and mechanisms of N entry, storage, and release from these pools have rarely been quantified. Furthermore, the fates of thatch and soil N pools following perturbations (e.g., turfgrass renovation practices) should be elucidated. The effect of 'catching or returning' clippings to turfgrass warrants additional study. Soil C sequestration can influence the magnitude of soil N storage under turfgrass; however the extent and rate of C sequestration, factors influencing C sequestration, and the relationship between soil C sequestration and N leaching needs further study (cf. Qian et al. 2003).

Improved knowledge of N and C cycles in turfgrass systems should guide the development of future management strategies aimed at minimising N leaching. Inexpensive and simple techniques for determining site-specific timings of fertiliser applications are needed by turfgrass managers. Research findings will only impact on N leaching in the field when management strategies developed for practitioners are shown to have minimal, if any, impact on turfgrass quality.

Conclusions

Turfgrass areas need not pose a risk to the environment if appropriate management strategies are undertaken. Irrigation scheduling that does not cause water to move beyond the rooting zone has been repeatedly shown to successfully decrease the amount of nitrate and ammonium leached from established turfgrass in temperate climates, without being detrimental to turfgrass growth or quality. The amount of water applied needs to match turfgrass requirements, but rates and frequencies should be chosen to avoid preferential flow. Irrigation scheduling can be optimised by using either soil sensor-controlled irrigation systems, or by calculating irrigation rates based on potential evapo-transpiration. Further strategies may be required to optimise irrigation regimes for turfgrass grown on coarse textured soils, such as use of soil amendments to decrease hydraulic conductivity and increase water-holding capacity, and in turn reduce leaching. Soil amendments that increase the CEC of coarse textured soils might decrease the leaching of ammonium, but not of nitrate.

Applying N fertilisers at rates and frequencies that match turfgrass requirements can also decrease N leaching from established turfgrass, although the benefits may be less marked under optimised irrigation regimes. Smaller, more frequent applications of water-soluble fertilisers reduce N leaching, as does the use of slow-release fertilisers. Turfgrass growth and quality are often also more consistent when water-soluble fertilisers are applied sparingly and frequently. Healthy turfgrass free from other nutrient deficiencies and free of disease and pests, should also ensure that uptake of applied N is maximised. Turfgrass appears to be more susceptible to N leaching during site establishment, and optimising both irrigation and fertiliser regimes while the turfgrass develops roots is particularly important for reducing N leaching.

Acknowledgements

Our present research on turfgrass at UWA is supported by Horticulture Australia Ltd (Project TU00007), WA Water Corporation, WA Waters & Rivers Commission, Scotts Australia, CRESCO/CSBP, Organic 2000, Turf Growers Association of WA, Golf Course Superintendents of WA, MicroControl Engineering (Rainman), City of Stirling, and City of Nedlands.

References

- Addiscott TM (1990). In 'Nitrate, agriculture, water: problems and challengers' (Ed. R Calvert), pp. 157–168. (INRA, Paris).
- Addiscott TM (1996). Measuring and modelling nitrogen leaching: parallel problems. *Plant and Soil* 181, 1–6.
- Adriano DC, Page AL, Elseewi AA, Chang AC and Staughan I (1980). Utilization and disposal of fly ash and other coal residues in terrestrial ecosystems: a review. *Journal of Environmental Quality* 9, 333–344.
- Adriano DC and Weber JT (2001). Influence of fly ash on soil physical properties and turfgrass establishment. *Journal of Environmental Quality* 30, 596–601.
- Amberger A (1989). Research on dicyandiamide as a nitrification inhibitor and future outlook. *Communications in Soil Science and Plant Analysis* 20, 1933–1955.
- Bauters TWJ, DiCarlo DA, Steenhuis TS and Parlange J (1998). Preferential flow in water-repellent sands. *Soil Science Society of America Journal* 62, 1185–1190.
- Beard JB (1973). 'Turfgrass: Science and culture'. (Prentice Hall, New York).
- Bigelow CA, Bowman DC and Cassel DK (2001). Nitrogen leaching in sand-based rootzones amended with inorganic soil amendments and sphagnum peat. *Journal of the American Society of Horticultural Science* 126, 151–156.
- Bowman DC, Cherney CT and Ruffy TW (2002). Fate and transport of nitrogen applied to six warm-season turfgrasses. *Crop Science* 42, 833–841.

- Bowman DC, Devitt DA, Engelke MC and Rufty TW (1998). Root architecture affects nitrate leaching from Bentgrass turf. *Crop Science* 38, 1633–1639.
- Brown KW, Duble RL and Thomas JC (1977). Influence of management and season on fate of N applied to golf greens. *Agronomy Journal* 69, 667–671.
- Bushoven JT, Jiang Z, Ford HJ, Sawyer CD, Hull RJ and Amador JA (2000). Stabilization of soil nitrate by reseeding with perennial ryegrass following sudden death. *Journal of Environmental Quality* 29, 1657–1661.
- Cameron KC, Smith NP, McLay CDA, Fraser PM, McPherson RJ, Harrison DF and Harbottle P (1992). Lysimeters without edge flow: an improved design and sampling procedure. *Soil Science Society of America Journal* 56, 1625–1628.
- Carpenter SR, Caraco NF, Correll DL, Howarth RW, Sharpley AN and Smith VH (1998). Nonpoint pollution of surface waters with phosphorus and nitrogen. *Ecological Applications* 8, 559–568.
- Carrow R (2004). Can we maintain turf to customers' satisfaction with less water? Proceedings of the 4th International Crop Science Congress, Brisbane. www.regional.org.au/au/cs.
- Engelsjord ME and Singh BR (1997). Effects of slow-release fertilizers on growth and on uptake and leaching of nutrients in Kentucky bluegrass turfs established on sand-based root zones. *Canadian Journal of Plant Science* 77, 433–444.
- Erickson JE, Cisar JL, Volin JC and Snyder GH (2001). Comparing nitrogen runoff and leaching between newly established St. Augustinegrass turf and an alternative residential landscape. *Crop Science* 41, 1889–1895.
- Fox RH and Bandel VA (1989). Dicyandiamide (DCD) research in agriculture in the mid-Atlantic region. *Communications in Soil Science and Plant Analysis* 20, 1957–1968.
- Geron CA, Danneberger TK, Traina SJ, Logan TJ and Street JR (1993). The effects of establishment methods and fertilization practices on nitrate leaching from turfgrass. *Journal of Environmental Quality* 22, 119–125.
- Gold AJ, DeRagon WR, Sullivan WM and Lemunyon JL (1990). Nitrate-nitrogen losses to groundwater from rural and suburban land uses. *Journal of soil and water conservation* 45, 305–310.
- Handson PD and Shelley BC (1993). A review of plant analysis in Australia. *Australian Journal of Experimental Agriculture* 33, 1029–1038.
- Hartley W (1950). The global distribution of tribes of the gramineae in relation to historical and environmental factors. *Australia Journal of Agricultural Research* 1, 355–373.
- Heckman JR, Liu H, Hill W, Demilia M and Anastasia WL (2000). Kentucky bluegrass responses to mowing practice and nitrogen fertility management. *Journal of Sustainable Agriculture* 15, 25–33.
- Hood EW, Williams MW and Caine N (2003). Landscape controls of organic and inorganic nitrogen leaching across an alpine/subalpine ecotone, Green Lakes Valley, Colorado Front Range. *Ecosystems* 6, 31–45.
- Huang ZT and Petrovic AM (1994). Clinoptilolite zeolite influence on nitrate leaching and nitrogen use efficiency in simulated sand based golf greens. *Journal of Environmental Quality* 23, 1190–1194.
- Hull RJ (1992). In 'Turfgrass' (Ed. DV Wadlington RN Carrow and RC Shearman), pp. 175–206. (American Society of Agronomy, Crop Science Society of America, Soil Science Society of America, Madison).
- Jiang Z, Bushoven JT, Ford HJ, Sawyer CD, Amador JA and Hull RJ (2000a). Mobility of soil nitrogen and microbial response following the sudden death of established turf. *Journal of Environmental Quality* 29, 1625–1631.
- Jiang Z and Hull RJ (1998). Interrelationships of nitrate uptake, nitrate reductase, and nitrogen use efficiency in selected Kentucky bluegrass cultivars. *Crop Science* 38, 1623–1632.
- Jiang Z, Sullivan WM and Hull RJ (2000b). Nitrate uptake and nitrogen use efficiency by Kentucky bluegrass cultivars. *HortScience* 35, 1350–1354.
- Kopp KI and Guillard K (2002). Clipping management and nitrogen fertilization of turfgrass: growth, nitrogen utilization, and quality. *Crop Science* 42, 1225–1231.

- Lewis DC, Grant IL and Maier NA (1993). Factors affecting the interpretation and adoption of plant analysis services. *Australian Journal of Experimental Agriculture* 33, 1053–1066.
- Liu H, Hull RJ and Duff DT (1993). Comparing cultivars of three cool-season turfgrasses for nitrate uptake kinetics and N recovery in the field. *International Turfgrass Society Research Journal* 7, 546–552.
- Liu H, Hull RJ and Duff DT (1997). Comparing cultivars of three-season turfgrasses for soil water NO₃⁻ concentration and leaching potential. *Crop Science* 37, 526–534.
- Macdonald AJ, Powlson DS, Poulton PR and Jenkinson DS (1989). Unused fertiliser nitrogen in arable soils—its contribution to nitrate leaching. *Journal of Science, Food and Agriculture* 46, 407–419.
- McLeod M, Aislabie J, Smith J, Fraser R, Roberts A and Taylor M (2001). Viral and chemical movement through contrasting soils. *Journal of Environmental Quality* 30, 2134–2140.
- McLeod M, Schipper LA and Taylor MD (1998). Preferential flow in a well drained and a poorly drained soil under different overhead irrigation regimes. *Soil Use and Management* 14, 96–100.
- Miltner ED, Branham BE, Paul EA and Rieke PE (1996). Leaching and mass balance of ¹⁵N-labeled urea applied to Kentucky Bluegrass turf. *Crop Science* 36, 1427–1433.
- Morton TG, Gold AJ and Sullivan WM (1988). Influence of overwatering and fertilization on nitrogen losses from home lawns. *Journal of Environmental Quality* 17, 124–130.
- Mosdell DK, Daniel WH and Freeborg RP (1986). Evaluation of dicyandiamide-amended fertilizers on Kentucky bluegrass. *Agronomy Journal* 78, 801–806.
- Nektarios PA, Petrovic AM and Steenhuis TS (2002). Effect of surfactant on fingered flow in laboratory golf greens. *Soil Science* 167, 572–579.
- OECD (1982). 'Eutrophication of waters: monitoring, assessment and control'. (OECD, Paris).
- Pathan SM, Aylmore LAG and Colmer TD (2001). Fly ash amendment of sandy soil to improve water and nutrient use efficiency in turf culture. *International Turfgrass Society Research Journal* 9, 33–39.
- Pathan SM, Aylmore LAG and Colmer TD (2002). Reduced leaching of nitrate, ammonium, and phosphorus in a sandy soil by fly ash amendment. *Australia Journal of Soil Research* 40, 1201–1211.
- Pathan SM, Aylmore LAG and Colmer TD (2003). Properties of several fly ash materials in relation to use as soil amendments. *Journal of Environmental Quality* 32, 687–693.
- Petrovic AM (1990). The fate of nitrogenous fertilizers applied to turfgrass. *Journal of Environmental Quality* 19, 1–14.
- Powlson DS (1988). In 'Nitrogen efficiency in agricultural soils' (Ed. DS Jenkinson and KA Smith), pp. 231–245. (Elsevier Applied Science Publishers Ltd, Barking).
- Qian Y and Follett RF (2002). Assessing soil carbon sequestration in turfgrass systems using long-term soil testing data. *Agronomy Journal* 94, 930–935.
- Qian YL, Bandaranayake W, Parton WJ, Mecham B, Harivandi MA and Mosier AR (2003). Long-term effects of clipping and nitrogen management in turfgrass on soil organic carbon and nitrogen dynamics: the CENTURY model simulation. *Journal of Environmental Quality* 32, 1694–1700.
- Reuter DJ and Robinson JB (1997). 'Plant analysis: an interpretation manual'. (CSIRO Australia, Collingwood).
- Sanchez CA and Doerge TA (1999). Using nutrient uptake patterns to develop efficient nitrogen management strategies for vegetables. *HortTechnology* 9, 601–606.
- Shepherd MA, Stockdale EA, Powlson DS and Jarvis SC (1996). The influence of organic nitrogen mineralization on the management of agricultural systems in the UK. *Soil Use and Management* 12, 76–85.
- Singleton PL, McLay CDA and Barkle GF (2001). Nitrogen leaching from soil lysimeters irrigated with dairy shed effluent and having managed drainage. *Australia Journal of Soil Research* 39, 385–396.
- Smith FW and Loneragan JF (1997). In 'Plant analysis: an interpretation manual' (Ed. DJ Reuter and JB Robinson), pp. 1–34. (CSIRO Australia, Collingwood).
- Smith VH (1998). In 'Successes, limitations, and frontiers in ecosystem science' (Ed. ML Pace and PM Groffman), pp. (Springer-Verlag, New York).

- Snyder GH, Augustin BJ and Davidson JM (1984). Moisture sensor-controlled irrigation for reducing N leaching in Bermudagrass turf. *Agronomy Journal* 76, 964–969.
- Snyder GH, Burt EO and Davidson JM (1981). Nitrogen leaching in Bermudagrass turf: Effect of nitrogen sources and rates. *Proceedings of the 4th International Turfgrass Research Conferences*, University of Guelph. International Turfgrass Society.
- Spangenberg BG, Fermanian TW and Wehner DJ (1986). Evaluation of liquid-applied nitrogen fertilizers on Kentucky bluegrass turf. *Agronomy Journal* 78, 1002–1006.
- Starr JL and DeRoo HC (1981). The fate of nitrogen fertilizer applied to turfgrass. *Crop Science* 21, 531–536.
- Starrett SK, Christians NE and Austin TA (1995). Fate of amended urea in turfgrass biosystems. *Communications in Soil Science and Plant Analysis* 26, 1595–1606.
- Turner TR and Hummel NW (1992). In 'Turfgrass' (Ed. DV Waddington RN Carrow and RC Shearman), pp. 385–440. (American Society of Agronomy, Crop Science Society of America, Soil Science Society of America, Madison).
- Waddington DV, Landschoot PJ and N W Hummel J (1989). Response of Kentucky bluegrass turf to fertilizers containing dicyandiamide. *Communications in Soil Science and Plant Analysis* 20, 2149–2170.
- Wang FL and Bettany JR (1994). Organic and inorganic nitrogen leaching from incubated soils subjected to freeze-thaw and flooding conditions. *Canadian Journal of Soil Science* 74, 201–206.
- Whitmore AP, Bradbury NJ and Johnson PA (1992). Potential contribution of ploughed grassland to nitrate leaching. *Agriculture, Ecosystems and Environment* 39, 221–223.
- Wilkinson JF and Miller RH (1978). Investigation and treatment of localized dry spots on sand golf greens. *Agronomy Journal* 70, 299–304.
- Zartman RE and Bartsch RA (1990). Using surfactants to enhance drainage from a dewatered column. *Soil Science* 149, 52–55.